ESCAPEE ORNAMENTAL PLANTS: EFFECT ON SPECIES DIVERSITY IN THE ADJOINING VEGETATION OF THE PARKS AND GARDENS, UNIVERSITY OF LAGOS, NIGERIA

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ABSTRACT

This study identified invasive species amongst the escapee OPs and determined their effects in the adjoining vegetation (AV) of the Parks and Gardens, University of Lagos, Nigeria. Species enumeration was done across the AV and an uninvaded vegetation located within the garden using systematic sampling technique in two consecutive years during dry (D1, D2) and wet (W1, W2) seasons. Escapee OPs were identified and Relative Importance Value (RIV) of plants were determined. Community structure was established using Shannon-Wiener (H') and Equitability (J) indices. Comparison of species was carried out using Jaccard's similarity coefficient (SC_I) . Results obtained revealed more individuals in the dry seasons and more in wet seasons for invaded and uninvaded vegetation respectively. A total of 59 species in 36 families occurred in invaded vegetation and 67 species in 39 families in the uninvaded vegetation. Diffenbachia seguine (Jacq.) Schott was the most important species in the invaded AV in all four seasons with mean RIV of 19.5%. In the uninvaded vegetation, mean RIV were low, ranging between 0.21% - 3.41%. Species diversity indices, H' and J were 2.39 - 3.03 and 0.69-0.71 respectively for invaded vegetation and 3.94-4.04 and 0.97 in all seasons respectively for uninvaded vegetation. SC_J ranged between 40.0-60.0% in the invaded vegetation and 84.4-90.1% in the uninvaded vegetation. Twelve OPs escaped into the AV. Among these, only four; Diffenbachia seguine, Synogonium podophyllum Schott, Heliconia psittacorum L.f. and Epipremnum pinnatum (L.) Engl. were invasive. Their invasion resulted in low species diversity in the adjoining vegetation.

Keywords: Ornamental plants, Invasive plants, Relative Importance Value, Adjoining vegetation, Plant diversity

INTRODUCTION

Man-made introductions in new habitats are responsible for rapid change within the indigenous communities (Ridenour and Callaway, 2001; Razanajatovo *et al.*, 2015). Many exotic plant species worldwide introduced for commercial

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exploitation, economic reasons and for ornamental purpose have subsequently become noxious invaders (Mack, 2003). Thus, horticulture is a major pathway for the introduction of alien plants (Razanajatovo *et al.*, 2015). Sometimes, under the alien conditions or in new invaded ecosystems, such species become naturalized and expand over native ecosystems (Richardson, 1998; Pyšek *et al.* 2011). It is estimated that as many as 50% of invasive species in general can be classified as ecologically harmful, based on their actual impacts (Richardson *et al.*, 2000).

Plants introduced for ornamental purposes, but have become invasive in their new areas with attendant environmental problems have been widely reported (Akobundu and Agyakwa, 1998; Reichard, and White, 2001). Tithonia diversifolia (Hemsl.) A. Gray, commonly called Mexican sunflower and native to Central America, Mexico and Cuba (Royal Horticultural Society, 1956), probably introduced into Africa as an ornamental plant (Akobundu and Agyakwa, 1998) is invasive in parts of Kenya, Uganda, Tanzania (Tropical Biology Association, 2010) and in Nigeria (Ayeni et al., 1997), where T. diversifolia is a widespread species, having colonized roadsides, waste places, fallow land and disturbed open spaces like abandoned construction sites, displacing native species (Adebowale and Olorode, 2005). Eichhornia crassipes (Mart.) Solms (Water hyacinth), a native of the Amazon Basin, was brought to Africa, probably to decorate ornamental ponds with its attractive violet coloured flowers (Akobundu and Agyakwa, 1998). Water hyacinth can explode into a floating blanket, affecting shipping, reducing fish catches, hampering electricity generation and human health. Eichhornia crassipes is highly invasive in Lagos State, Nigeria (Adekanmbi et al., 2009).

Leucaena leucocephala (Lam.) de Wit (Fabaceae) native to Southern Mexico and Northern Central America (Germplasm Resources Information Network (GRIN), 1995), is a fast-growing, multipurpose and beneficial tree, suitable as an ornamental and roadside landscaping species (Orwa *et al.*, 2009). It has become an aggressive invader in tropical and subtropical disturbed areas. The plant is reported to be invasive in more than 20 of the more than 105 countries where it was introduced (Walton, 2003). *Dieffenbachia seguine*, an ornamental plant commonly used as an exotic houseplant, has its native range from the Caribbean to Tropical South America (Missouri Botanical Garden, 2016). The plant is reported to have escaped into the wild and has become invasive in many locations including Fiji Island (Smith, 1979), the Pacific Islands (Space and Flynn, 2000),

Kosrae Island, Federated States of Micronesia (Space *et al.*, 2000) and Nigeria (Bassey and Akpanumun, 2009; Aigbokhan, 2013).

Invasive alien species pose a threat to native plant communities globally, especially where these communities are disturbed (D'Antonio et al., 2001). Invasions by non-native plants can alter ecosystem functions and reduce native plant diversity (Levine et al., 2003). Plant invasions dramatically affect the distribution, abundance and reproduction of many native species (Sala et al., 1999). For example, the introduction of pines in the Southern Hemisphere has affected large areas of natural grass and shrub lands, bringing a lot of changes in the dominant life forms, decreasing the species composition and modifying vegetation patterns and nutrient cycles in the region (Richardson et al., 1994). A study of the impact of invasive plants on invaded communities revealed that species richness, diversity and evenness were reduced in invaded plots (Hejda et al., 2009). Invasive species exhibiting the greatest impact reduced species numbers per plot and the total number of species recorded in the communities sampled by almost 90%. Furthermore, a strong reduction of species number at the plot scale resulted in a marked reduction in the total species number at the landscape scale and in less similarity between invaded and uninvaded vegetation (Hejda et al., 2009).

Owing to the devastating impact of invasive plants in the ecosystem, and knowing that ornamental plant introductions are a major pathway of their inlet into a region, there is need for early detection of these species by studying the attributes of ornamental plants in places where they are introduced. Therefore, this study was conducted to identify invasive escapee ornamental plants and determine their effects in the adjoining vegetation of the Parks and Gardens, University of Lagos, Nigeria.

METHODOLOGY

Sites description

The Parks and Gardens, University of Lagos is situated in Akoka campus of the University of Lagos, Yaba, Lagos, southwestern Nigeria and is located 6° 30' 40.10" N, 3° 23' 54.39" E and 07.01 m above sea level (Figure 1). It was set up in 1981 to perform the main functions of developing and carrying out a landscaping plan for the University campus, perfect general beautification and environmental care of the campus, plant and maintain shade trees and flowers as well as fruit and tree nurseries both for the University and for sale to staff and the public. The

University of Lagos is largely surrounded by the Lagos lagoon and the vegetation in this area is represented by mangrove and riparian type (Nodza *et al.*, 2014). The cultivated garden covers an approximate total area of $4,551 \text{ m}^2$. It is adjoined by natural vegetation, which for most parts is waterlogged and experience seasonal flooding. This section is invaded by some ornamental plants. Lying at right angle to this area is well drained soil with rich uninvaded vegetation.



Figure 1: A Map showing the location of Parks and Gardens, University of Lagos

Sampling procedure

This study was conducted between 2013 and 2015. All ornamental plants grown in the garden (in the soil and in pots) were identified (Table 1). In addition, all herbaceous (weedy) plants in the graden floor were also identified and enumerated (Sanyaolu *et al.*, 2018a). Sampling of the the AV was by the belt transect technique (Grant *et al.*, 2004; Ahmed, 2016). A single transect was laid in each across the vegetation five meters away from the edge of the garden and was continuous through the entire length, running parallel to the edge of the garden. Sampling along the transect was at 10 meters intervals using a 2 m x 2 m wooden quadrat. All plants rooted within each quadrat were identified and enumerated. This procedure was also repeated in the uninvaded vegetation located within the garden premises. A total of eight and four quadrats respectively were taken in the invaded and uninvaded vegetation. Data were collected in dry and wet seasons of two consecutive years. These were denoted D1, D2, W1 and W2 for first dry, second dry, first wet and second wet seasons respectively. Identification of plants was done on-site using keys from standard floras and manuals. Samples of

unidentified plants were taken to the Herbarium located in the Department of Botany, Obafemi Awolowo University for identification.

Analysis of data

Data obtained for both invaded and uninvaded sections of the garden were analysed to determine Relative Importance Value (RIV) (Curtis, 1959; Barbour et al., 1999; Awodovin et al., 2013), plant diversity indices and Jaccard similarity coefficient as follows:

Relative Importance Value RIV (%) = [Relative density + Relative frequency]/2

Relative Density RD (%) = $[density of a species/density of all plant species] \times 100$. Eq 1

Species density $D(m^2) =$ Total no.of individuals of a species in all quadrats/total no.of quadrats taken

Relative frequency of a species RF(%) =[frequency of that species/frequency of all species] $\times 10$ Eq 3

.....Eq 2

Species frequency F (%) = No. of quadrats species occur/total no. of quadrats $\times 100$ Eq 4

Determination of plant community structure

Shannon-Wiener and Equitability indices were obtained using the Paleontological software package (PAST) version 3.10 (Hammer et al., 2015).

Shannon-Wiener Index $H' = -\sum (pi, \ln pi)$ (Shannon and. Wiener, 1949) Eq 5

Where pi = ni/N, where ni is the no. of individuals of a species in a garden and N is the total number of individuals in the garden.

 $V = H' / \ln S$ (Whittaker, 197Eq 6 Equitability index

Where, \overline{S} = total number of species enumerated in the garden and In = natural logarithm

Similarity index This was determined using jaccard's similarity coefficient SCI (Jaccard, 1912) and calculated using the formula: $SC_I = [w/(A + B - w)] \times 100$Eq 7

(Spellberg, 1991; Awodoyin *et al.*, 2013). Where $\overline{SC_J}$ = Jaccard similarity coefficient, w = number of species common to (or shared by) two seasons/plots, A = number of species in one season/plot and B = number of species in the second season/plot

RESULTS

Species enumeration in the adjoining vegetation

Results obtained for ornamental plants (Ops) composition in the garden unit of the study site revealed a total of 123 species disrtributed in 43 families, among which were the invasive species found in the AV (Table 1). Results for enumerated species revealed that more individual plants were enumerated in the invaded vegetation, occurring in the dry season, compared to the uninvaded vegetation and also where more individuals occurred in the wet season (Tables 1 and 2). Results for species composition of the invaded AV are presented in Tables 1. A total of 296, 297, 292 and 268 individuals were identified in the D1, D2, W1 and W2 respectively. These were distributed among 35 species belonging to 27 families, 42 species in 29 families, 31 species in 24 families and 32 species in 24 families respectively during these seasons (Tables 1). In the uninvaded vegetation, 113, 107, 133 and 129 individuals were enumerated for D1, D2, W1 and W2 respectively. These were distributed in 59, 59, 63 and 65 species and 35, 34, 37 and 37 families respectively for the seasons, signifying that although fewer individuals occurred in the uninvaded sections, these individuals were distributed among more plant species and families compared to the invaded vegetation (Table 2).

	Lagos		
S/ N	Family	S/N	Scientific name
1	Acanthaceae	1.	Crossandra infundibuliformis L.
1	Acaminaccae	1. 2.	Pseuderanthemum carruthersii (Seem.) Guillaumin
		2. 3.	Sanchezia speciosa Ruiz & Pav
2	Agavaceae	3. 4.	Agave marginata
2	Agavaceae	т . 5.	Yucca sp.
3	Amaranthaceae	5. 6.	Alternanthera dentata
3 4	Amaryllidaceae	0. 7.	Crinum jagus (J.Thomps.) Dandy
-	Amarymuaeeae	8.	Crinum mooreii 'alba'
		0. 9.	Crinum xanthophyllum
). 10.	Hymenocallis latifolia (Mill.) M. Roem
		10.	Zephyranthes rosea Lindl.
5	Annonaceae	11.	Greenwayodendron suaveolens (Engl. & Diels) Verdc
5	Annonaceae	12.	Polyalthia longifolia (Sonn.) Thwaites
6	Apocynaceae	13.	Allamanda cathartica L
0	ripocynaceae	14.	Amsonia sp
		15. 16.	Catharanthus roseus (L.) G.Don
		10.	Plumeria rubra L.
		17.	Rauvolfia serpentina (L.) Benth. ex Kurz
		10. 19.	Tabernaemontana coronaria
7	Araceae	20.	Aglaonema commutatum Schott
,	Thueede	20.	Caladium bicolor (Aiton) Vent.
		21.	Asterix Dieffenbachia seguine (Jacq.) Schott
		22.	Asterix Epipremnum pinnatum (L.) Engl.
		24.	Peltandra virginica (L.) Schott
		25.	Philodendron atabapoense G.S.Bunting
		26.	Philodendron hederaceum (Jacq.) Schott
		27.	Asterix Syngonium podophyllum Schott
8	Araliaceae	28.	Polyscias batfluoriana (Sander ex André) L.H. Bailey
0	1 11 11 11 10 0 10 0	29.	Polyscias filicifolia (C.Moore ex E.Fourn.) L.H.Bailey
		30.	Polyscias scutellaria (Burm.f.) Fosberg
		31.	Schefflera arboricola (Hayata) Merr.
9	Arecacea	32.	Roystonea regia (Kunth) O.F.Cook
	1 11000000	33.	Adonidia merrillii (Becc.) Bec
		34.	Archontophoenix alexandrae. (F. Muell.) H. Wendl. & Drude
		35.	<i>Caryota</i> sp
		36.	Dypsis lutescens (H.Wendl.) Beentje & J.Dransf
		37.	Licuala sp.
		38.	Pritchardia remota Becc.
10	Asparagaceae	39.	Asparagus densiflorus 'sprengeri'
	ParaBarrao	40.	Aspidistra elatior Blume
		41.	Chlorophytum bichetii

 Table 1: Ornamental plant composition of the Parks and Gardens, University of Lagos

	42.	Chlorophytum comosum (Thunb.) Jacques
	43.	Cordyline fruticosa (L.) A. Chev
	44.	Dracaena marginata Lam
	45.	Dracaena fragrans (L.) Ker Gawl
	46.	Dracaena loureiri Gagnep.
	47.	Dracaena surculosa Lindl.
	48.	Furcraea foetida (L.) Haw
	49.	Sansevieria cylindrica Bojer ex Hook.
	50.	Sansevieria liberica Gérôme & Labroy
	51.	Sansevieria trifasciataPrain.
	52.	Sansevieria zeylanica (L.) Willd
Cactaceae	53.	Gymnocalycium sp.
	54.	Pereskia bleo (Kunth) DC.
Cannaceae	55.	Canna indica L.
Combretaceae	56.	Terminalia catappa L
	57.	Terminalia mantaly H. Perrier
Commelinaceae	58.	Tradescantia spathacea Sw.
	59.	Tradescantia zebrine hort. ex Bosse
Costaceae	60.	Costus woodsonii Maas
Crassulaceae	61.	Bryophyllum pinnatum (Lam.) Oken.
Cupressaceae	62.	Thuja occidentalis L.
Cycadaceae	63.	Cycas petrea
	64.	<i>Cycas revoluta</i> Thunb.
	65.	Acalypha hispida Burm.f.
Euphorbiaceae	66.	Acalypha wilkesiana Müll. Arg.
	67.	Breynia disticha J.R. Forst. & G. Forst.
	68.	Codiaeum variegatum (L.) A. Juss
	69.	Euphorbia ingens E.Mey. ex Boiss.
	70.	Euphorbia milii L.
	71.	Jatropha curcas L.
	72.	Jatropha integerrima. Jacq.
	73.	Jatropha multifidi L.
Fabaceae	74.	Delonix regia (Bojer ex Hook.) Raf
	75.	Poinciana pulcherrima L. (Caesalpinia pulcherrima (L.)

75. $\frac{1000}{\text{Sw.}}$

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- 76. Senna alata (L.) Roxb.
- 77. Senna corymbosa (Lam.) Irwin & Barneby
- Gentianaceae 78. Anthocleista nobilis G. Don
- Heliconiaceae 79. Asterix Heliconia psittacorum L. f.
- Lythraceae 80. *Cuphea hyssopifolia* Kunth.
 - 81. Lagastroemia indica L.
 - 82. Lagerstroemia speciosa (L.) Pers
- Malvaceae 83. *Hibiscus rosa-sinensis* L.
 - 84. *Sterculia tragacantha* Lindl.
 - Marantaceae 85. *Calathea majestica*
 - 86. Calathea orbifolia (Linden) H.Kenn
 - 87. *Ctenanthe setosa* (Roscoe) Eichler

26 27 28	Moraceae Moringaceae Myrtaceae	 88. 89. 90. 91. 92. 93. 94. 95. 96. 97. 	Thaumatococcus daniellii (Benn.) Benth. Ficus benjamina L Ficus elastica Roxb. ex Hornem. Ficus lyrata Warb Ficus microcarpa L. Ficus pumila L. Moringa oleifera Lam Eucalyptus sp. Eugenia uniflora L. Syzygium jambos L. (Alston)
29	Nephrolepidaceae	98. 99.	Nephrolepis exaltata (L.) Schott Nephrolepis falcata (Cav.) C. Chr.
30	Nyctaginaceae	100.	Bougainvillea sp.
31	Pandanaceae	101.	Pandanus odoratissimus L.
32	Phytolaccaceae	102.	Petiveria alliacea L.
33	Pinaceae	103.	Pinus Sp.
34	Poaceae	104.	Axonopus compressus (Swartz) P. Beauv
		105.	Cymbopogon citratus (DC. ex Nees)
35	Polypodiaceae	106.	Phymatosorus sp
36	Rubiaceae	107.	Gardenia jasminoides
		108.	Gardenia jasminoides variegata
		109.	Hamelia patens Jacq.
		110.	Ixora coccinea L
		111.	Ixora finlaysoniana Wall. ex G.Don
		112.	Mussaenda philippica A. Rich.
		113.	Pseudomussaenda flava Verdc.
37	Rutaceaae	114.	Murraya paniculata (L.) Jack
38	Solanaceae	115.	Brunfelsia pauciflora (Cham. & Schltdl.) Benth.
		116.	Solanum torvum Sw.
39	Strelitziaceae	117.	Ravenola madagascariensis Sonn
30	Turnaraceae	118.	Turnera ulmifolia L
31	Verbenaceae	119.	Duranta erecta L.
		120.	Duranta erecta' variegata'
42	Xanthorrhoeaceae	121.	Aloe vera (L.) Burm.f.
43	Zingiberaceae	122.	<i>Cucurma</i> sp.
		123.	Ethligera eliator (Jack) R.M. Sm.

*Ornamental plants that invaded the adjoining vegetation of the garden

Table 2:	Species composition of invaded Adjoining vegetation of the Parks and
	Gardens, University of Lagos in two wet and dry seasons

		RIV (%)					
Family	Scientific name	D1	D2	W1	W2	Mean	
1.Acanthaceae	1. Asystasia gangetica (L.) T.Anderson	0.00	0.00	1.04	2.61	0.91	
2. Amaranthaceae (2)	2. Alternanthera brasiliana (L.) Kuntze	0.00	0.73	0.00	0.00	0.18	
	3. Cyathula prostrata (L.) Blume	10.23	6.21	7.17	2.70	6.58	
 Amaryllidaceae 	4. Crinum jagus (J.Thomps.) Dandy	2.93	0.00	0.00	0.00	0.73	
4. Annonaceae	5. Greenwayodendron suaveolens (Engl. & Diels) Verdc	0.00	0.00	0.78	0.00	0.20	
	6. Alstonia boonei De Wild.	0.87	0.00	0.00	0.84	0.43	
5. Apocynaceae (3)	7. Alstonia congensis Engl.	0.87	0.73	2.32	0.00	0.98	
··· • • • • • • • • • • (•)	8. Holarrhena floribunda (G.Don) T.Durand & Schinz	0.87	0.00	0.00	0.00	0.23	
	9. Rauvolfia vomitoria Afzel	0.00	2.17	1.04	1.68	1.22	
6. Araceae (6)	10. Caladium bicolor (Aiton) Vent.	0.00	0.00	0.00	0.84	0.21	
	11. Asterix Dieffenbachia seguine (Jacq.) Schott	17.36	14.11	22.54	24.04	19.51	
	12. Epipremnum pinnatum (L.) Engl.)	3.76	6.09	4.89	3.35	4.52	
	13. Peltandra virginica (L.) Schott	3.27	5.54	0.91	2.05	2.94	
	14. Philodendron atabapoense G.S.Bunting	0.00	1.40	0.00	0.00	0.35	
	15. Asterix synogonium podophyllum Schott	10.57	2.58	4.89	4.09	5.53	
7. Arecaceae	16. Elaeis guineensis Jacq.	2.93	5.48	0.78	1.68	2.72	
3. Asparagaceae	17. Dracaena arborea (Willd.) Link	0.87	0.89	1.56	0.00	0.83	
9. Bignoniaceae (2)	18. Crescentia cujete L.	0.00	1.45	0.00	0.00	0.36	
0 ()	19. Newbouldia laevis (P. Beauv.) Seem. ex Bureau	2.93	2.29	1.56	2.89	2.42	
10. Bixaceae	20. Bixa orellana L	0.87	1.95	0.00	0.00	0.71	
11. Combretaceae	21. Terminalia catappa L	0.87	1.79	0.00	0.00	0.67	
12. Commelinaceae (2)	22. Aneilema aequinoctiale P.Beauv.	0.00	1.40	0.00	0.00	0.35	
	23. Commelina erecta L	0.00	0.00	0.00	0.84	0.21	
13. Connaraceae (2)	24. <i>Byrsocarpus coccineus</i> Thonn. ex Schumach.	0.00	0.73	1.56	0.00	0.57	
	25. Cnestis ferruginea Vahl ex DC	0.00	0.00	1.04	0.00	0.26	
14. Convolvulaceae	26. Ipomoea involucrata P. Beauv.	0.00	0.73	0.00	0.00	0.18	
15. Dioscoreaceae (2)	27. Dioscorea bulbiferia L	0.00	0.73	1.69	3.34	1.44	
()	28. Dioscorea dumetorum (Kunth) Pax	0.00	0.00	0.00	0.84	0.21	
16. Ebanaceae	29. Diospyros mobutensis Gurke	0.00	0.73	0.00	0.00	0.18	
17. Euphorbiaceae	30. Alchornea cordifolia (Schumach. & Thonn.) Müll.Arg.	3.46	4.35	3.89	8.10	4.95	
18. Fabaceae (3)	31. <i>Albizia ferruginea</i> (Guill. & Perr.) Benth.	0.00	0.73	0.00	0.00	0.18	
	32. Albizia lebbeck (L.) Benth.	0.87	0.73	0.00	0.84	0.61	
	33. Leptoderris micrantha Dunn	0.00	0.73	0.00	0.00	0.18	
19. Gentianaceae	34. Anthocleista nobilis G. Don	0.87	2.17	0.78	2.05	1.47	
20. Heliconiaceae	35. Asterix heliconia psittacorum L. f.	4.58	1.74	6.16	5.03	4.38	
	36. Harungana madagascariensis Poir.	3.44	7.48	4.52	4.93	5.09	
21. Hypericaceae							

23. Lecythidaceae	38. Napoleonaea vogelii Hook. & Planch	0.87	0.73	2.32	0.84	1.19
24. Malvaceae	39. <i>Glyphaea brevis</i> (Spreng.) Monachino	0.00	1.79	0.00	0.00	0.45
25. Menispermaceae	40. Cissampelos owariensis Beauv. ex DC	0.87	1.45	0.78	0.84	0.99
26. Moraceae (6)	41. Antiaris toxicaria (Pers.) Lesch.	0.87	1.57	0.00	1.68	1.03
	42. Trilepisium madagascariense DC. (Bosqueia angolensis Ficalho)	0.00	0.73	0.00	0.00	0.18
	43. Ficus capensis Thunb.	0.87	1.4	1.56	2.51	1.59
	44. Ficus exaasperat Roxb. ex Hornem.	0.00	0.00	0.78	0.00	0.20
	45. <i>Ficus vogelii</i> (Miq.) Miq	0.00	0.73	0.00	0.84	0.39
27. Musaceae	46. <i>Musa</i> sp.	2.59	0.00	2.34	1.86	1.70
28. Myrtaceae	 Syzygium samarangense (Blume) Merr. & L.M.Perry 	0.87	0.00	0.00	0.00	0.22
29. Passifloraceae	48. Adenia lobata (Jacq.) Engl	2.07	1.45	5.81	1.68	2.75
30. Phyllanthaceae	49. Phyllanthus niruri L	0.00	0.00	0.00	0.84	0.21
31. Poaceae	50. Bambusa vulgaris Schrad. ex J.C. Wendl.	6.99	5.14	8.99	9.40	7.63
32. Polygonaceae	51. Antigonon leptopus Hook. & Arn	0.00	0.00	0.00	0.84	0.21
33. Rubiaceae (2)	52. Chassalia kolly (Schumach.) Hepper	1.73	1.23	0.00	0.00	0.74
	53. Leptodermis sp.	0.87	0.00	0.00	0.00	0.22
34. Sapindaceae (2)	54. Blighia sapida K.D.Koenig	0.87	0.89	0.00	0.00	0.44
	55. Paullinia pinnata L	2.41	2.90	2.97	2.51	2.70
35. Vitaceae (3)	56. Cissus aralioides (Welw. ex Baker) Planch	0.00	0.00	0.78	0.00	0.20
	57. Cissus populnea Guill.& Perr	0.00	0.00	0.00	0.84	0.21
	58. Cissus quadrangularis L.	0.87	1.45	0.78	0.00	0.78
36. Zingiberaceae	59. Etlingera elatior (Jack) R.M. Sm	1.04	1.40	1.56	0.00	1.00
		100.2	100.1	100.1	100.1	99.8
	Total individual plants enumerated	296	297	292	268	
	No of species	35	42	31	32	
	Families	27	29	24	24	
D1: 1 st dry season	D2: 2 nd dry season W1	: 1 st wet	season V	V2: 2 nd wet	t season	

*Invasive ornamental plants with high mean RIV in the adjoining vegetation

Table 3:	Species	composition	of	uninvaded	vegetation	of	the	Parks	and
Gardens, Uni	versity of	Lagos in two	wet	and dry sea	sons				

/	V	<u> </u>	RIV (%)				
Family	S/N	Scientific name	D1	D2	W1	W2	RIV
1. Acanthaceae (1)		Asystasia gangetica (L.)					
		T.Anderson	1.96	2.36	2.78	2.08	2.30
2. Amaranthaceae (2)		Alternanthera brasiliana (L.)					
		Kuntze	1.96	2.87	1.60	1.69	2.03
		Cyathula prostrata (L.) Blume	2.93	1.40	2.4	2.93	2.42
3. Anacardiaceae (1)		Mangifera indica L.	1.96	0.96	1.18	1.69	1.45
4. Apocynaceae (4)		Alstonia boonei De Wild.	0.98	0.96	0.80	0.85	0.90
		Alstonia congensis Engl.	1.96	1.40	2.40	2.08	1.96
		Holarrhena floribunda (G.Don)					
		T.Durand & Schinz	1.96	1.92	1.60	1.69	1.80
		Rauvolfia vomitoria Afzel	1.45	1.40	1.60	2.08	1.63
5. Araceae (2)		Caladium bicolor (Aiton) Vent.	1.96	2.36	1.98	1.69	2.00
		Peltandra virginica (L.) Schott	0.00	0.96	0.80	0.85	0.65
6. Arecaceae (2)		Cocos nucifera L.	0.98	0.96	0.80	0.85	0.90
		Elaeis guineensis Jacq.	2.93	2.87	1.60	1.69	2.27
7 Asteraceae (1)		Emilia praetermissa Milne-Redh	0.00	0.00	0.00	1.24	0.31
8. Bignonaceae (2)		Newbouldia laevis (P. Beauv.)					
•		Seem. ex Bureau	3.40	2.87	2.40	2.54	2.80
		Tabebuia rosea (Bertol) DC	0.98	0.00	0.80	0.85	0.66
9. Bixaceae (1)		Bixa orellana L	0.98	0.00	0.80	0.85	0.66
10. Combretaceae (2)		Terminalia catappa L	2.42	3.12	2.40	2.08	2.51
		Terminalia superba	0.98	0.96	1.60	0.85	1.10
11. Commelinaceae (2)		Aneilema aequinoctiale P.Beauv.	1.96	2.36	3.18	1.69	2.30
		Commelina erecta L	1.96	0.96	2.78	2.54	2.06
12. Connaraceae (2)		Byrsocarpus coccineus Thonn.					
		ex Schumach.	0.98	0.96	0.80	0.85	0.90
		Cnestis ferruginea Vahl ex DC	1.96	1.92	1.60	1.69	1.79
 Convolvulaceae (1) 		Ipomoea involucrata P. Beauv.	0.00	0.96	1.60	0.00	0.64
14. Cyperaceae (2)		Cyperus tenuis Sw.	2.42	2.36	1.98	2.93	2.42
		Cyperus cyperoides	0.98	0.96	0.80	0.85	0.90
15. Dioscoreaceae (2)		Dioscorea bulbiferia L	1.96	0.96	1.60	1.24	1.44
		Dioscorea dumetorum (Kunth)					
		Pax	0.98	0.96	0.80	0.85	0.90
16. Ebanaceae (1)		Diospyros mobutensis Gurke	1.96	0.96	1.60	1.24	1.44
17. Euphorbiaceae (1)		Alchornea cordifolia (Schumach.					
		& Thonn.) Müll.Arg.	0.98	1.92	0.80	0.85	1.14
18. Fabaceae (3)		Albizia ferruginea (Guill. & Perr.)					
		Benth.	1.45	0.96	1.18	1.24	1.21
		Albizia lebbeck (L.) Benth.	0.98	0.96	0.80	0.85	0.90
		Leptoderris micrantha Dunn	0.00	1.92	1.98	0.85	1.19
19 Gentianaceae (1)		Anthocleista nobilis G. Don	0.98	1.92	1.60	0.85	1.34
20. Hypericaceae (1)		Harungana madagascariensis	• • •	•	•	•	.
.		Poir.	3.40	3.76	3.53	2.93	3.41
21. Icacinaceae (1)		Icacina trichantha Oliver.	3.40	3.32	2.40	2.93	3.01
22. Irvinginaceae (1)		Irvinginia gabonensis	0.98	0.96	0.80	0.85	0.90
23. Lauraceae (1)		Presea Americana Mill	0.98	0.96	0.80	0.00	0.69

24. Lecythidaceae (1)	<i>Napoleonaea vogelii</i> Hook. & Planch	1.96	1.92	1.60	1.69	1.79
25. Malvaceae (2)	Glyphaea brevis (Spreng.)	1.00	1.02	1.00	1.00	1.70
(_)	Monachino	0.98	0.96	0.80	0.85	0.90
	Sida acuta Burm. f.	1.96	0.00	1.98	2.54	1.62
26. Meleaceae (1)	Azadirachta indica	0.98	3.32	2.78	2.08	2.29
27. Menispermaceae (1)	Cissampelos owariensis Beauv.					
	ex DC	0.98	0.00	1.60	1.69	1.07
28. Moraceae (7)	Antiaris toxicaria (Pers.) Lesch.	2.93	3.32	3.58	2.54	3.09
	Artocarpus heterophyllus	0.98	0.96	1.22	1.69	1.21
	Trilepisium madagascariense					
	DC.	1.96	1.92	1.60	2.08	1.89
	Ficus capensis Thunb.	1.45	1.40	0.80	1.69	1.34
	Ficus exasperata Roxb. ex			0.00		
	Hornem.	1.96	1.92	1.60	1.69	1.79
	Ficus Thonningii Blume	1.96	1.92	0.00	0.85	1.18
	Ficus vogelii (Miq.) Miq	1.96	1.92	1.60	1.69	1.79
29. Musaceae (1)	Musa sp.	2.42	2.36	1.60	2.54	2.23
30. Myrtaceae (1)	Syzygium samarangense					
	(Blume) Merr. & L.M.Perry	0.98	0.96	0.80	0.85	0.90
31. Passifloraceae (1)	Adenia lobata (Jacq.) Engl	0.98	0.96	1.60	0.85	1.10
32. Phyllanthaceae (2)	Phyllanthus niruri L	0.98	0.96	1.60	1.24	1.20
	Phyllanthus amarus Schum &	0.00	0.00			
	Thonn.	0.00	0.96	0.00	2.47	0.86
33. Poaceae (2)	Bambusa vulgaris Schrad. ex	0.00	0.00	0.00		0.00
	J.C. Wendl.	2.93	2.87	1.18	2.08	2.27
	Setaria barbata (Lam.) Kunth	0.00	0.96	1.60	0.85	0.85
34. Polygonaceae (1)	Antigonon leptopus Hook. & Arn	1.96	1.92	1.60	0.85	1.58
35. Pyroiodea (1)	Morinda lucida Benth	0.00	0.00	0.80	0.85	0.41
36. Rubiaceae (3)	Chassalia kolly (Schumach.)	0.00	0.00	0.00	0.00	••••
	Hepper	2.93	2.36	2.40	2.93	2.66
	Leptodermis sp.	1.96	1.92	2.40	1.69	1.99
	Oldenlandia corymbosa (L.)	1.96	1.92	1.60	2.08	1.89
37. Sapindaceae (2)	Blighia sapida K.D.Koenig	0.98	0.96	0.80	0.85	0.90
	Paullinia pinnata L	0.98	0.96	2.40	0.85	1.30
38. Vitaceae (3)	Cissus aralioides (Welw. ex					
	Baker) Planch	0.00	0.00	0.00	0.85	0.21
	Cissus populnea Guill.& Perr	0.98	0.00	0.80	0.85	0.66
	Cissus quadrangularis L.	0.98	0.96	0.80	0.85	0.90
39. Zingiberaceae (1)	Etlingera elatior (Jack) R.M. Sm	0.98	1.92	0.80	1.69	1.35
J		100.		100.	100.	
		2	99.9	1	1	100.1
	Total individuals	113	107	133	129	
	No of species	59	59	63	65	
	Families	35	34	37	37	

Escapee and invasive ornamental plants identified in the adjoining vegetation

Table 1 shows the Relative importance values of all plants enumerated in the AV during the study period. *Dieffenbachia seguine*, an introduced ornamental plant grown in the garden, remained the most important species in all four seasons of study with mean RIV of 19.5% (Table 1). Result also revealed that *D. seguine* had higher RIV in the wet seasons (22.5% and 24.0% for W1 and W2 respectively), when the total number of species appeared lower compared to dry season (17.4% and 14.1% for D2 and D2 respectively), when number of species were higher. Highest RIV for *D. seguine* (24.0%) occurred in the last season of study (W2).

Twelve ornamental plants grown in the garden (Table 1) escaped into the AV. These were *Crinum jagus*, *Dieffenbachia seguine*, *Epipremnum pinnatum*, *Peltandra virginica*, *Caladium bicolor*, *Synogonium podophyllum*, *Heliconia psittacorum*, *Polyalthia suaveolens* (formerly *Greenwayodendron suaveolens*) *Dracaena arbore*, *Syzygium samarangense*, *Philodendron atabapoense* and *Terminalia catappa*. Of these, only *Dieffenbachia seguine* (Araceae), *Synogonium podophyllum* (Araceae), *Epipremnum pinnatum* (Araceae) *and Heliconia psittacorum* (Heliconiaceae), (Plate 1 A, B. C and D respectively) with mean RIV 19.51%, 5.53%, 4.52% and 4.38% respectively, were invasive the AV. Five of the garden ornamental plants were found in the uninvaded vegetation located within the garden premises. These include *Peltandra virginica*, *Caladium bicolor*, *Syzygium samarangense*, *Terminalia catappa* and *Synogonium podophyllum*. However, they only occurred in insignificant numbers compared to other species in the vegetation, thus showing no invasion (Table 2).

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Plate 1: Invasive ornamental plants in the study site (all arrowed); *Epipremnum pinnatum* as a climber and creeper (A) *Heliconia psittacorum* (B) *Dieffenbachia seguine* (C) and *Synogonium podophyllum* (D)

Plant species diversity and similarity indices

Results obtained revealed that species diversity were higher in uninvaded vegetation compared to invaded vegetation. The Result as presented in Table 3 shows that species lower diversity and similarity in the invaded vegetation compared to the uninvaded vegetation. In the invaded vegetation, species diversity indices were generally higher in the dry compared to the wet seasons. Shannon diversity index ranged between 2.39 and 3.03, the highesr values occurring in the dry seasons. Equitability index ranged between 0.69 and 0.81, with the higher values also observes in the dry seasons (Table 4). In the uninvaded vegetation, species diversity indices were higher in the wet compared to the dry seasons (Table 2). Shannon diversity index ranged between 3.94 - 4.04, the higher values occurring in the wet seasons, whereas, equitability was 0.97 in all four seasons (Table 3). Similarity index between vegetation of the study sites

is shown in Table 4. In the invaded section of the AV, lowest similarity was 40% occurring between D1 and W1 whereas, highest similarity values of 60% occurred between W1 and W2. In the uninvaded section of the garden, similarity index ranged between 84.4% and 91.0%, with highest values also occurring between W1 and W2. Similarity index between invaded and uninvaded sections of AV was 61.5%.

 Table 4:
 Diversity index of species in four seasons in invaded and uninvaded sections of the AV

	Invadeo	section			Uninv	aded sectior	۱	
Diversity index\Season	D1	D2	W1	W2	D1	D2	W1	W2
Shannon	2.54	3.03	2.39	2.41	3.97	3.94	4.02	4.04
Equitability	0.73	0.81	0.70	0.69	0.97	0.97	0.97	0.97
D1-First dry season	W1_ Fi	rst wet se	ason D	2- second d	Irv	$W^2 - Seco$	ond wet	season

D1-First dry season W1- First wet season D2- second dry W2 – Second wet season

Table 5:	Jaccard's similarity index (%) comparing species in four seasons in the
	study site

Seasons	Invade	d section	1		Uninva	ion		
	D1	D2	W1	W2	D1	D2	W1	W2
D1	-				-			
D2	50.98				84.38	-		
W1	40.00	46.00			90.63	87.69	-	
W2	45.65	45.10	60.00	-	90.77	85.07	91.04	-
								-

D1-First dry season W1- First wet season D2- second dry W2 - Second wet season

DISCUSSION

The University of Lagos is largely surrounded by the Lagos lagoon (Nodza *et al.*, 2014) and the study site, the Parks and Gardens of the University, is located by a swamp that leaves a larger section of the AV soil waterlogged, particularly during the wet seasons. Since most plants require a well drained soil for optimal growth and reproduction, this may have partially accounted for lower species number in the wet season in the AV, when only water-logged soil tolerant plants could thrive. However, during the dry season, the adjoining soil conditions improved, allowing more species to emerge, hence, higher number of species recorded in dry seasons. This seasonal fluctuation in soil water content may have also created the necessary disturbance required by invasive escapees to thrive (D'Antonio *et al.*, 2001). Thus, moist soil conditions and seasonal flooding as well as fewer plants in the wet seasons in the AV may have presented the opportunity for aggressive invasion by these species. However, in the uninvaded vegetation, relatively higher individual plants, species number and families were enumerated in the wet compared to dry seasons as is generally observed in well drained tropical soils

(Peters *et al.*, 2014). The vegetation in this area was observed to be very stable with regards to the number and type of species as indicated by higher similarity index between species in the four seasons of study. It could be inferred that there is a direct correlation between soil moisture content and type and abundance of plant species in the vegetation as earlier reported by Peters *et al.* (2014).

Although many OPs escaped into the adjoining vegetation, only a few of these species were invasive. This validates earlier report that only a few among the species that escaped into the landscapes may become invasive (Masterson, 2007). Fountain (2016) stated some characteristic of an invasive plant to include; plants that may have ornamental characteristics making it highly desirable to gardeners and enhances spread to new areas, adaptation to a variety of cultural conditions, very rapid growth, reproduction by vegetative means, flowers and fruits at a young age, produces large numbers of seeds and seeds to have a high percent germination or the species is already known to be invasive in other regions. All or some of these characteristics were observed in each of the species listed as invasive escapees in this study. All four invasive escapees were listed among the commonly grown non native OPs in Southwestern Nigeria (Sanyaolu *et al.*, 2018b). Therefore, their widespread usage in the region may further enhance their spread into other native landscapes, where they may continue their invasions, thus posing serious threat to the region.

Invasion or potential invasion by all four species has been previously reported. Dieffenbachia seguine, a plant native to North and South America (Cuartas-Hernandez and Nunez-Farfan, 2006), the most invasive species in the AV, has been reported to be invasive in many locations especially in the Pacific Islands (American Samoa) and elsewhere in Nigeria. In Akwa Ibom and Delta states of Nigeria, D. segune forms extensive mats and appear to be displacing native and naturalized flora at the location (Bassey and Akpanumun, 2009). In this study, invasion by these ornamental plants was largely restricted to the section of the AV with moist, water logged soils. This agrees with previous reports that D. seguine can be invasive and difficult to remove in moist areas (Space and Imada, 2004; Bassey and Akpanumun, 2009). Moreover, in this study, the RIV of D. segune, was highest in the last season of study compared to the previous seasons, signifying increased population as time progressed. This indicates a gradual invasion, which may probably be accompanied by gradual elimination of native plants. This agrees with earlier report that D. seguine can thrive in dense shade, crowding out other species (Space and Imada, 2004). Aigbokhanei (2013) stated that D. seguine infested sites were relatively sparsely and poorly vegetated when compared with adjoining non-infested sites, signifying that the plant probably was out-competing native plants with which it exists in the community.

Epipremnum pinnatum (commonly called pothos), native to Mo'orea, French Polynesia is listed as invasive in the Galápagos Islands, Tanzania, St. Lucia and islands in the Pacific including Hawaii, French Polynesia, and Micronesi (Graveson, 2012). *Syngonium podophyllum* indigenous to South and Central America from Mexico to Panama (Global Invasive Species Database (GISD), 2015), has established invasive populations in the United States, South Africa, Singapore, the Caribbean, and on several Pacific islands, where it establishes dense populations that displace native plants and grow over native trees and surrounding vegetation (Ferriter *et al.*, 2005; Morgan and Overholt, 2005; GISD, 2010. *Heliconia* species native to South America (National Tropical Botanical Garden (NTBG), 2016), is reported to have escaped from gardens, becoming naturalized along roadsides in Hawaii, although it is yet to be listed as serious invasive species.

Invasions by non-native plants can reduce native plant diversity (Sala et al., 1999; Levine et al., 2003; Hejda et al., 2009), species richness and evenness (Hejda et al., 2009) in invaded plots. These reports validate the results obtained from the present study where plant diversity indices appeared low in all seasons, with much lower values in wet seasons, when D. seguine invasion was particularly highest. In addition, species diversity values were higher in invaded vegetation of the AV compared to nearby uninvaded vegetation. Higher similarity index between species in the wet seasons in the invaded vegetation was probably due to large population of D. seguine and low species diversity recorded during these seasons. Overall, species diversity in the unvaded vegetation remained higher in all season compared to invaded vegetation indicating stability in the uninvaded vegetation. Thus, this study confirms earlier report that the conventional approach of introduction and landscaping with exotic ornamental plants in urban centres may affects biodiversity, as it limits the diversity of native species in areas dominated by these plants and can affect surrounding natural environments (Hostetler and Main, 2010).

CONCLUSION

Data from the present study revealed that twelve ornamental plants cultivated in the garden unit escaped into the adjoining vegetation of the garden. Four of these; *Dieffenbachia seguine, Synogonium podophyllum, Heliconia psittacorum* and *Epipremnum pinnatum* were invasive and their invasion resulted in low species

diversity in invaded compared to uninvaded vegetation. Findings from this study bring to the fore the need to regulate the introduction of exotic ornamental plants into the country. Development as well as use of native ornamental plant is therefore advocated.

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